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Integrated Bayesian Multi-model Approach to Quantify Input, Parameter and Conceptual Model Structure Uncertainty in Groundwater Modeling

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1 Highlights

2	٠	Full Bayesian multi-model approach to quantify uncertainty of MODFLOW model
3	٠	Simultaneously quantifies model structure, input and parameter uncertainty
4	٠	DREAM with a novel likelihood function is combined with BMA
5	٠	Neglecting conceptual model uncertainty results in unreliable prediction
6	٠	Results in more reliable model predictions and accurate uncertainty bounds
7		

8 Abstract

9 A flexible Integrated Bayesian Multi-model Uncertainty Estimation Framework (IBMUEF) is presented to simultaneously quantify conceptual model structure, input and parameter 10 uncertainty of a groundwater flow model. In this fully Bayesian framework, the DiffeRential 11 Evolution Adaptive Metropolis (DREAM) algorithm with a novel likelihood function is 12 combined with Bayesian Model Averaging (BMA). Four alternative conceptual models, 13 representing different geological representations of an overexploited aquifer, have been 14 developed. The uncertainty of the input of the model is represented by multipliers. A novel 15 likelihood function based on a new heteroscedastic error model is included to extend the 16 applicability of the framework. The results of the study confirm that neglecting conceptual 17

18 model structure uncertainty results in unreliable prediction. Consideration of both model 19 structure and input uncertainty are important to obtain confident parameter sets and better 20 model predictions. This study shows that the IBMUEF provides more reliable model 21 predictions and accurate uncertainty bounds.

Keywords: Conceptual model structure uncertainty, Bayesian approach, Input
uncertainty, Bayesian model averaging, Uncertainty quantification, Groundwater flow
model.

25 **1. Introduction**

26 The reliability of predictions of numerical groundwater flow models is strongly influenced by 27 different sources of uncertainty. To ensure reliable predictions and decision support in sustainable water resources management, it is important to assess all different sources of 28 29 uncertainty. Conceptual model structure uncertainty can be related to the complexity of a 30 groundwater model (Elshall and Tsai, 2014), which may vary from a simple to a detailed representation of the processes and geological information of the groundwater system (Rojas 31 32 et al., 2010; Mustafa et al., 2019). The geological structure is often very complex and 33 heterogeneous and only partially known. Hence, the incomplete and biased representation of the processes, and the complex structure of a system often result in uncertainty in model 34 predictions (Refsgaard et al., 2006; Rojas et al., 2008). 35

It is important to assess the different sources of uncertainty to ensure accurate predictions and 36 reliable decision support in sustainable water resources management. The conventional 37 treatment of uncertainty in groundwater modeling primarily focuses on parameter 38 uncertainty, whereas uncertainties due to the model structure are often neglected (Gaganis & 39 Smith, 2006; Rojas et al., 2008). However, many researchers have recently acknowledged 40 that the uncertainty arising from the conceptual model structure has a significant effect on the 41 model predictions and that parameter uncertainty does not cover the whole range of 42 uncertainty (Bredehoeft, 2005; Højberg & Refsgaard, 2005; Mustafa et al., 2018, 2019; 43 Neuman, 2003; Poeter & Anderson, 2005; Refsgaard et al., 2006, 2007; Rojas et al., 2008; 44 45 Troldborg et al., 2007). Therefore, neglecting conceptual model structure uncertainty may result in unreliable predictions and underestimation of the total predictive uncertainty. 46

47 Most of recent studies only consider a single conceptual model structure and may fail to48 adequately sample the relevant space of plausible conceptual models. Single model

49 techniques are unable to account for errors in model output resulting from structural 50 deficiencies of a specific model as single models cannot capture all hydrogeological 51 processes of the system (Ajami et al., 2007; Rojas et al., 2008; Mustafa et al., 2019). As a 52 consequence, a well-calibrated model does not always accurately predict the behavior of the 53 dynamic system (Van Straten & Keesman, 1991). Choosing a single model out of equally 54 plausible alternative models may contribute to either type I (reject true model) or type II (fail 55 to reject false model) model errors (Li & Tsai, 2009; Neuman, 2003).

56 Bredehoeft (2005) has presented different examples where the collection of new data and unforeseen elements challenged well-established conceptual models. Hence, researchers in 57 hydrogeological science have suggested to use different alternative conceptual models for a 58 single hydrogeological system (Højberg & Refsgaard, 2005; Mustafa et al., 2019; Nettasana 59 60 et al., 2012; Refsgaard et al., 2006; Troldborg et al., 2007). Such multi-model approaches can be used to estimate a broader uncertainty band so that it is more likely to include the 61 unknown true predicted value (Rojas et al., 2010). However, conceptual model structure 62 uncertainty arising from the simplified representation of the hydro(geo)logic processes, 63 geological stratification and boundary conditions, has received less attention (Refsgaard et 64 65 al., 2006; Rojas et al., 2010).

A model averaging technique can be used to combine predictions of multiple models. 66 Hydrologists have been using different model averaging techniques to obtain an average 67 prediction and a reliable uncertainty band from a number of plausible conceptual models 68 (Vrugt, 2016a). The predictions of multiple models are combined by using weights, which 69 70 can be equal or can be determined through regression-based approaches (Yin and Tsai, 2018). 71 Poeter & Anderson (2005) have proposed an approach in which weights are connected to model performance and the predictions of the conceptual models are combined using 72 73 Akaike's weights (Akaike, 1974). However, in the multi-model predictions, this approach 74 does not consistently include prior knowledge about parameters and conceptual models. Refsgaard et al. (2006) have proposed a method to incorporate prior knowledge of multiple 75 model structures. In this approach, a set of conceptual models are calibrated separately and 76 the consistency of these models was assessed using pedigree analysis. However, this method 77 does not provide results in a quantitative way that can be used to analyse uncertainty in terms 78 79 of probabilities.

80 On the other hand, the Bayesian Model Averaging (BMA) method (Draper, 1994; Hoeting et al., 1999) derives predictions from a set of alternative conceptual models to construct a 81 predictive uncertainty distribution using probabilistic techniques. The weights in the BMA 82 method are assessed based on the relative performance of each model to reproduce system 83 behavior during the observation period. Recently, BMA has received attention of researchers 84 in diverse fields because of its more reliable and accurate predictions than other existing 85 model averaging methods (Li & Tsai, 2009; Rojas et al., 2008, 2010; Singh et al., 2010; 86 Troldborg et al., 2010; Vrugt, 2016a; Ye et al., 2004, 2010). 87

88 An important challenge in implementing Bayesian Model Averaging is evaluating Bayesian model evidence (BME). There are different techniques to evaluate BME, such as analytical 89 techniques, mathematical approximations, and numerical evaluation. The analytical solution 90 91 is strongly depended on the assumptions. That is why exact and computationally efficient analytical solutions are rarely available (Schoniger et at., 2014). There are different methods 92 93 of mathematical approximation, such as Laplace approximation, Kashyap Information criterion, Bayesian Information Criterion and Akaike Information Criterion. Those different 94 mathematical information criterions may provide contradictory results in model ranking and 95 posterior model weights (Poeter and Anderson, 2005; Singh et al., 2010; Ye et al., 2010; 96 Schoniger et al. 2014). However, awareness about the contradictory results from different 97 methods is very limited (Hoge et al., 2019). Although numerical methods are as prone to be 98 biased than mathematical approximations, Schoniger et al. (2014) have concluded that bias-99 free numerical evaluation methods are better than mathematical approximations for model 100 selection. Among the numerical evaluation methods, the multi-chain Markov Chain Monte 101 Carlo (MCMC) based DiffeRential Evolution Adaptive Metropolis (DREAM) algorithm 102 became very popular because of its statistical robustness and numerical efficiency (Leta et al., 103 104 2015; Vrugt et al., 2008, 2016; Laloy et al., 2013;). However, applications of this algorithm for quantifying conceptual structural uncertainty of a real-world groundwater flow model also 105 considering uncertainties from the model input and parameters are very limited. 106

Maximum Likelihood Bayesian Model Averaging (MLBMA), which is an approximation of BMA, has been applied recently in hydrogeology to analyse the predictive distribution of several conceptual models (Neuman, 2003; Ye et al., 2004). MLBMA depends on the calibration of alternative conceptual model parameters. However, by using this method estimated biased parameters will compensate conceptual model structure errors during calibration to obtain the best model fit (Højberg & Refsgaard, 2005; Refsgaard et al., 2006; 113 Troldborg et al., 2007). Refsgaard et al. (2006) have reported that the model becomes biased 114 when calibrated models are used for simulating variables that were not included in 115 calibration.

However, the existing Bayesian averaging approach does not quantify the uncertainty arising from the different components of the individual conceptual model and how they affect the model prediction (Tsai, 2010; Gupta et al., 2012; Tsai and Elshall, 2013). Tsai and Elshall (2013) and Chitsazan and Tsai (2015) address this issue by introducing the Hierarchical BMA (HBMA) method. In this HBMA method, the uncertainty arising from the different components of the individual conceptual model is considered using a BMA tree.

122 Alternative approaches to account for conceptual model structure uncertainty along with uncertainty from other sources are integrated uncertainty assessment approaches, which 123 124 combine estimation of individual sources of uncertainty into an integrated modeling framework. In surface water hydrology, two distinct approaches have been developed and 125 applied: Bayesian total error analysis (BATEA) (Kavetski et al., 2006a, 2006b; Kuczera et 126 al., 2006) and the integrated Bayesian uncertainty estimator (IBUNE) (Ajami et al., 2007). 127 Both methods consider model parameter, input and conceptual structural uncertainties to 128 quantify model prediction uncertainties. However, model ranking or multi-model 129 combinations are not considered in the BATEA framework. Hence, diagnostic model 130 comparison is not possible in this framework. On the other hand, the IBUNE framework 131 allows to combine multi-model predictions based on model weights obtained from a non-132 Bayesian optimization algorithm. As a consequence, a robust Bayesian derivation of posterior 133 probabilities is missing. To quantify input uncertainties, the IBUNE framework uses a 134 135 multiplier that is assumed to be independent and normally distributed with fixed mean and variance. Vrugt and Robinson (2007) have criticized this assumption as it is not a very 136 137 appropriate way to quantify model input and conceptual structural uncertainties. Furthermore, identification of spatial and temporal variation of the input multipliers is not possible in this 138 framework as it considers only a single multiplier. The latter might result in a biased 139 estimation of input uncertainties and thereby result in biased predictive uncertainty. As 140 141 groundwater model input data, such as recharge and abstraction rates, are usually estimated using indirect methods or specific models which are not accurate and can present errors both 142 143 in space and time, the IBUNE approach is often not suitable for groundwater modeling.

In the field of groundwater hydrology, however, no systematic integrated framework has 144 been proposed to date. Rojas et al. (2008) have applied BMA in combination with the 145 generalized likelihood uncertainty estimation (GLUE) method (Beven, 1993; Beven & 146 Binley, 1992) to quantify conceptual model structure uncertainty. A three-dimensional 147 hypothetical setup with three alternative conceptualizations has been considered to 148 demonstrate the method. However, some researchers have criticized GLUE because it is not a 149 formal Bayesian method and may result in statistically incoherent and unreliable parameters 150 and predictive distributions (Mantovan & Todini, 2006; Montanari, 2005; Stedinger et al., 151 152 2008). Therefore, the likelihood and threshold used for model selection and weighting in the approach of Rojas et al. (2008) has a lack of statistical basis and, as a consequence, 153 conceptual model structure and parameters are not optimized in this method, which could 154 result in overestimation of predictive uncertainty (Nettasana et al., 2012). 155

Recently, Xue & Zhang (2014) have applied multimodel ensemble Kalman filter (EnKF) in 156 157 combination with the Bayesian model averaging framework to explicitly consider the model structural uncertainty. They advocated that the EnKF is computationally more efficient 158 compared to other existing Bayesian methods. However, uncertainty arising from model 159 input and measurement heteroscedasticity has not been explicitly considered in this 160 framework. The performance of this multimodel EnKF framework has been tested using 161 synthetic 2D conceptual groundwater model in idealized conditions without consideration of 162 observational uncertainty or model bias, whereas the real-world models are often three-163 dimensional and more complex, and observations are not bias free (Hoge et al. 2019). Ridler 164 165 et al. (2018) have also criticized this multimodel EnKF framework because of its limitation in application with bias observation. Hendricks Franssen et al. (2011) reported that the EnKF 166 significantly outperformed with synthetic experimental data compare the real data. 167

168 Mustafa et al. (2018) presented a Bayesian approach to simultaneously quantify parameter and input uncertainty of a groundwater flow model. The performance of this approach has 169 170 been evaluated using a single conceptual real-world groundwater flow model. Groundwater recharge and groundwater abstraction multipliers with a spatial and temporal character have 171 been introduced in this study to quantify the uncertainty of the spatially distributed input data 172 of the groundwater model along with parameter uncertainty. Nevertheless, the conceptual 173 174 model structural uncertainty has not been considered in this study. As a result, the latter study is unable to account for the errors in the model output resulting from the structural 175 176 deficiencies. Recently, Mustafa et al. (2019) presented a multi-model approach to quantify

groundwater-level prediction uncertainty considering alternative conceptual models. In the 177 second study, the combined effect of conceptual model structure, the climate change and 178 groundwater abstraction scenarios on future groundwater-level prediction uncertainty has 179 been evaluated. However, alternative conceptual models of this study have been calibrated 180 using a local optimization method and considering only model parameter. As a result, this 181 approach is unable to account for the uncertainty arising from the model input and 182 parameters. Estimated biased parameters will compensate conceptual model structural errors 183 during calibration to obtain the best model fit, as it relies on a single optimum parameter set. 184 185 Moreover, the approach is missing the statistical robustness because of its deterministic modelling approach. 186

Very recently, Hoge et al. (2019) highlight the difference between BMA and Bayesian 187 188 combined model averaging (BCMA) following Minka (2002) and Monteith et al. (2011). According to Hoge et al. (2019), BCMA means the application of equations for BMA 189 190 (section 2.3) to forecast combinations of individual conceptual models instead of the application of equations for BMA to the individual conceptual model alternatives. Hoge et al. 191 (2019) concluded that the objective of the modelling should be the main driver in selecting 192 model averaging approaches. They also suggested to use BCMA instead of BMA if the 193 objective of the modelling is to increase the reliability of the model prediction. The Integrated 194 Bayesian Uncertainty Estimator (IBUNE) that has been applied in surface water hydrology 195 by Ajami et al. (2007) has been considered as a practical application of applying BMA in 196 similar fashion of BCMA (Hoge et al. 2019). However, as mentioned earlier, Ajami et al. 197 (2007) allows to combine multi-model predictions based on model weights obtained from a 198 non-Bayesian optimization algorithm. As a consequence, a robust Bayesian derivation of 199 posterior probabilities is missing. 200

Hence, more research on a systematic integrated fully Bayesian framework is needed to quantify the uncertainty arising from the conceptual model structure, inputs and parameters of groundwater flow models with consideration of the heteroscedasticity of the groundwater level error. Additionally, the application of such an integrated multimodel framework on real-world cases is necessary to better understand the impacts of different sources of uncertainty on real-world model calibration and prediction problems.

The general objective of this study is therefore the development and application of an
Integrated Bayesian Multi-model Uncertainty Estimation Framework (IBMUEF) to quantify

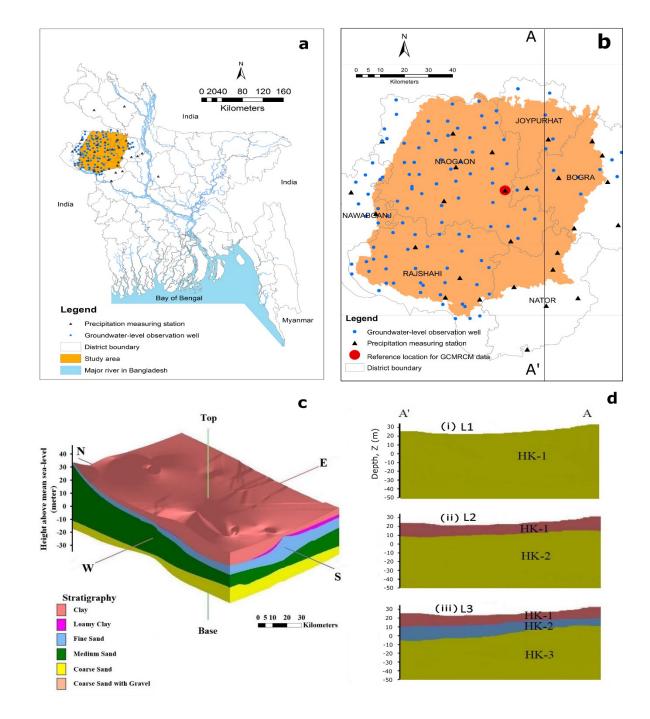
input, parameter, measurement and conceptual model structure uncertainty of a fully 209 distributed physically-based groundwater flow model to provide reliable predictions of 210 groundwater system. In the proposed integrated fully Bayesian multi-model framework, the 211 DiffeRential Evolution Adaptive Metropolis (DREAM) algorithm with a specific likelihood 212 function is combined with the Bayesian Model Averaging (BMA) framework. In this new 213 DREAM-BMA methodology, a likelihood function has been included based on the novel 214 heteroscedastic error model for groundwater levels proposed by Mustafa et al. (2018). Like 215 IBUNE of Ajami et al. (2007), the current study uses equations for BMA in a similar fashion 216 217 as BCMA. However, unlike Ajami et al. (2007), our study allows to combine multi-model predictions based on model weights obtained from a Bayesian optimization algorithm. This is 218 the first attempt to apply a fully Bayesian multi-model framework in real-world groundwater 219 modeling to quantify conceptual model structure uncertainty along with uncertainties 220 originating from model input, parameters and measurement error. In this methodology, the 221 fully Bayesian approach proposed by Mustafa et al. (2018) has been combined with the 222 Bayesian Combined Model Averaging (BCMA) to simultaneously quantify the uncertainty 223 arising from the conceptual model structural, input and parameter of a fully distributed 224 groundwater flow model. Additionally, the proposed approach is applicable for all types of 225 226 residual errors i. e. both for homoscedastic and heteroscedastic errors. The IBMUEF is a flexible framework as (i) there is no limitation for the number or complexity of alternative 227 conceptual models, (ii) users can choose the number and dimensions (spatial and temporal) of 228 input multipliers, (iii) both quantitative and qualitative information of the system can be used 229 230 in the alternative conceptual models, and (iv) it is applicable for both homoscedastic and heteroscedastic residuals errors. Moreover, the proposed approach is able to avoid 231 232 compensation for conceptual model structural uncertainty arising from biased parameter estimates obtained from a model fit, as it does not rely on a single optimum parameter set. 233

Finally, the framework (IBMUEF) is applied in an over-exploited aquifer in the northwestern Bangladesh, as it is necessary to understand the impacts of conceptual model structural uncertainties on model prediction in realistic conditions. The specific objectives of this paper are: (i) to quantify model uncertainty originating from errors in model conceptualization, (ii) to quantify individual uncertainty contributions arising from model input, parameter, and measurement and conceptual model uncertainties, (iii) to understand conceptual model structure uncertainty impacts on calibration and model prediction, (iv) to evaluate the applicability of our approach for groundwater models in realistic conditionsusing alternative conceptual groundwater flow models.

243 **2. Methodology**

244 **2.1 Study area**

The study area covers the six north-western districts of Bangladesh (Figure 1a). The aquifer 245 consists mainly of medium sand, coarse sand and coarse sand with gravel, with minor 246 fractions of clay, loamy clay, and fine sand (Figure 1c). The thickness of each stratigraphic 247 unit moreover varies spatially. The average thickness of the top layer is 18 m and it consists 248 of clay, clayey loam and fine sand. A 20 m thick medium sand layer is present below the top 249 layer. The bottom part of the aquifer consists of a 35 m thick layer of coarse sand and coarse 250 sand with gravel. Average rainfall is between 1400 mm and 1550 mm per year. However, 251 rainfall distribution is not uniform over the year. There is almost no rainfall during the dry 252 season (November to April), which is the major cropping season in this study area (Mustafa 253 et al., 2017b). The area is mainly covered by irrigated agriculture of which more than 80 % is 254 rice. Irrigated agriculture uses around 97 % of total groundwater abstraction (Shahid, 2009; 255 Mustafa et al. 2017a). Groundwater level in this study area is continuously decreasing due to 256 overexploitation of groundwater for irrigation (Mustafa et al., 2017a). 257



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Figure 1: Description of the study area: (a) Location of the study area in the north-western part of Bangladesh; (b) study area with precipitation measurement stations (triangles) and groundwater observation wells (circles); (c) stratigraphy of the study area; (d) cross-sectional (A-A') view of different hydrogeological models: (i) one-layered model (L1), (ii) two-layered model (L2), (iii) three-layered model (L3). Taken from Mustafa et al. (2019).

266 **2.2 Bayesian approach to quantify input and parameter uncertainty**

Mustafa et al. (2018) presented a Bayesian approach to simultaneously quantify parameter and input uncertainty of a fully distributed groundwater flow model. For the details of the approach we refer the reader to Mustafa et al. (2018). A short summary of the approach is presented here. A hydrogeological model can be defined as follows:

$$0 = M\left(\bar{I}, \theta, \eta\right) \tag{1}$$

Where \bar{I} and O represent the input and output matrix of model M; θ and η are the parameters and boundary conditions of the corresponding model. To quantify input uncertainty along with parameter uncertainty, following Kavetski et al. (2002, 2006a, 2006b) a modified concept of multipliers for a fully distributed groundwater model has been introduced by Mustafa et al. (2018). The uncertainty of the input data (groundwater abstraction and recharge) is quantified using the following input error model:

$$I_{ij} = \bar{I}_{ij} * m_{ij} \tag{2}$$

Where $\bar{I}_{ij} = {\bar{\iota}_{1,1}, \bar{\iota}_{1,2}, \bar{\iota}_{1,3}, ..., ..., \bar{\iota}_{J,N}}$ represents the initial input for the ith month and jth location, m_{ij} is the respective input multiplier and I_{ij} represents the corresponding corrected input. m_R represents the groundwater recharge multipliers while m_A represents groundwater abstraction multipliers (Table 1). The multipliers are considered as an additional individual latent parameter and are estimated along with the model parameters.

Traditionally, residual errors in groundwater modelling are considered to be homoscedastic. However, Mustafa et al. (2018) have shown that the standard deviation of the groundwater level residual is not always constant but may increase with the deviation of groundwater level from the normal. In this study, the long-term average is considered as the normal groundwater level. A novel heteroscedastic error model for groundwater level has been proposed in this fully Bayesian approach to consider the heteroscedasticity of the groundwater level residual. The proposed heteroscedastic error model is defined as follows:

$$\sigma = A * |SH_i - \overline{OH}| + B \tag{3}$$

289

Where σ is standard deviation, A is a parameter representing the groundwater level uncertainty slope, B is a parameter representing the groundwater level uncertainty intercept, SH_i represents the simulated groundwater level for each time step and \overline{OH} represents the observed long-term (30 years for this study) average groundwater level. The log-likelihood function proposed by Vrugt et al. (2009a, 2013) has been adopted and modified by Mustafa et al. (2018) for spatially distributed groundwater models. The proposed novel heteroscedastic error model for groundwater level has been incorporated in this modified log-likelihood function. The new log-likelihood function is as follows:

$$\ell(\theta|\bar{I},\bar{O},\eta) = -\frac{T}{2}\ln(2\pi) - \sum_{l=1}^{L} \left(\sum_{t=1}^{T} \ln(\sigma_{tl})\right) - \frac{1}{2}\sum_{l=1}^{L} \left(\sum_{t=1}^{T} \left(\left(\frac{\bar{O}_{tl} - O_{tl}}{\sigma_{tl}}\right)^2\right)\right)$$
(4)

Where $\overline{O} = \{\overline{o}_1, \overline{o}_2, \overline{o}_3, ..., ..., \overline{o}_T\}$ represents the output series of observed groundwater levels in observation wells, $O = \{o_1, o_2, o_3, ..., ..., o_T\}$ represents the output series of simulated groundwater levels for the same observation well, $t = \{1, 2, 3, ..., ..., T\}$ represents time step, T represents the total number of time steps, $l = \{1, 2, 3, ..., ..., ..., L\}$ represents the location of the observation wells and L represents the total number of observation wells.

This log-likelihood function has been used in this study because of (i) its numerical stability, 304 (ii) algebraic simplicity and (iii) its applicability for both homoscedastic and heteroscedastic 305 residual errors. To sample the posterior distribution based on the likelihood function 306 (Equation 4), the DREAM-ZS sampler has been used. The Differential Evolution Adaptive 307 Metropolis algorithm (DREAM) is a multi-chain Markov Chain Monte Carlo (MCMC) 308 simulation algorithm introduced by Vrugt et al. (2008; 2009a; 2009b). The DREAM-ZS 309 algorithm (Vrugt, 2016) has been used in this study to explicitly quantify the uncertainty 310 311 arising from model input and parameters of a groundwater flow model. More details about 312 the DREAM algorithm are explained in Vrugt et al. (2008; 2009a; 2009b) and Vrugt (2016).

In this study, we extend this approach to include conceptual model structure uncertainties and we improve the methodology by combining it with Bayesian Model Averaging (BMA).

315 2.3 Bayesian Model Averaging (BMA)

Bayesian Model Averaging is a probabilistic scheme for combining predictions from multiple conceptual models to provide a more realistic and reliable description of total prediction uncertainty. It is a technique that can be used to account for model structural uncertainty (Madigan et al., 1996). It is a statistical procedure that derives average predictions by weighing predictions from different models in such a way that the weighted prediction is a better representation of the observed system variables compared to any individual model of the ensemble. The BMA prediction gives higher weights to better performing models, as the agreement between the model predictions and the observations is assumed to be a measure of the model likelihood. The variance of BMA is a measure of the uncertainty of BMA prediction. The variance of BMA predictions is representing both the within-model variance and the between-model variance.

Bayesian Model Averaging (BMA) has been used to deduce more reliable predictions of groundwater levels than the predictions produced by the different individual groundwater models. Draper (1994) and Hoeting et al. (1999) present an extensive overview of BMA. Here, only a short summary of BMA is given.

Consider $\mathbf{M} = [M_1, M_2, M_3, ..., M_K]$ the set of alternative conceptual models, $\mathbf{Y} = \{y_1, y_2, ..., y_n\}$ is a 1 × n observation vector of a quantity of interest, F_{jk} is the point forecast of each alternative conceptual model for $j = \{1, 2, ..., n\}$ observations and $k = \{1, 2, ..., K\}$ models. Now by combining the different conceptual models forecasts in a matrix **F** having dimensions of n × K, the weighted average forecast of the quantity of interest is

$$y_j = \sum_{k=1}^{K} \beta_k F_{jk} + e_j \tag{5}$$

Where $\beta = \{\beta_1, \beta_2, \dots, \beta_K\}$ represents the weight vector of each conceptual model and e_j is noise.

As we know, model predictions are associated with uncertainty. The uncertainty can be described using a probability density function (forecast distribution) p(.). When applying BMA, assuming uniform prior distribution the posterior predictive distribution of the quantity of interest is given by

$$p(y_j|F_{jk}) = \sum_{k=1}^{K} p(y_j|F_{jk}, M_k) p(M_k|F_{jk})$$
(6)

Where, p(.|.) = conditional probability density function (PDF), $p(y_j|F_{jk}, M_k) =$ posterior predictive distribution of y_j on F_{jk} under the considered model M_k and $p(M_k|F_{jk}) =$ posterior probability of the respective model M_k . This is also known as the likelihood (weight) of the corrected model M_k .

The BMA predictive mean and variance of y are conditional to the discrete ensemble of the proposed alternative conceptual models, M (Draper, 1994).

$$E[y_{j}|F_{jk}] = E_{M}[E(y_{j}|F_{jk}, M)] = \sum_{k=1}^{K} E[y_{j}|F_{jk}, M_{k}] p(M_{k}|F_{jk})$$
(7)

$$Var[y_{j}|F_{jk}] = E_{M}[Var(y_{j}|F_{jk}, M)] + Var_{M}[E(y_{j}|F_{jk}, M)]$$

= $\sum_{k=1}^{K} Var[y_{j}|F_{jk}, M_{k}] p(M_{k}|F_{jk}) + \sum_{k=1}^{K} (E[y_{j}|F_{jk}, M_{k}] - E[y_{j}|F_{jk}])^{2} p(M_{k}|F_{jk})$ (8)

349

Where $E[y_j|F_{jk}, M_k]$ and $Var[y_j|F_{jk}, M_k]$ represent, respectively, the expected value and variance of y_j on F_{jk} under the considered conceptual model, M_k . Considering $E[y_j|F_{jk}, M_k] = y_k$, $Var[y_j|F_{jk}, M_k] = \sigma_k^2$ and $p(M_k|F_{jk}) = \beta_k$, the BMA predictive mean and variance of the quantity of interest can be developed as follows

$$E[y_j|F_{jk}] = \sum_{k=1}^{K} y_k \beta_k$$
(9)

354

$$Var[y_{j}|F_{jk}] = \sum_{k=1}^{K} \sigma_{k}^{2} \beta_{k} + \sum_{k=1}^{K} \beta_{k} \left(y_{k} - \sum_{u=1}^{K} y_{u} \beta_{u} \right)^{2}$$
(10)

The first term of the variance is representing the within-model variance, while the second term represents the between-model variance.

The BMA method considers the uncertainty of each model's forecast and uses it to develop a 357 predictive distribution rather than only a weighted average. So, the BMA method provides an 358 average forecast along with an associated forecast distribution. The forecast distribution can 359 be used for constructing confidence intervals. This BMA forecast density enforces one 360 significant constraint for the weights, i.e., $\beta_k \ge 0$ and $\sum_{k=1}^{K} \beta_k = 1$ to avoid the development of 361 unrealistic forecast distributions (e.g., densities can even become negative without this 362 restriction). For successful application of the BMA method, proper estimates of the weights, 363 and standard deviation, of the normal conditional pdfs of the ensemble members are needed. 364 365 To estimate the weights and standard deviation, the log-likelihood function is used for algebraic simplicity and numerical stability, 366

$$\mathcal{L}(\beta_{BMA}, \sigma_{BMA} | \mathbf{F}, \mathbf{Y}) = \sum_{j=1}^{n} \log \left\{ \sum_{k=1}^{K} \beta_k \frac{1}{\sqrt{2\pi\sigma_k^2}} \exp\left[-\frac{1}{2}\sigma_k^{-2} (y_j - F_{jk})^2\right] \right\}$$
(11)

367 where β_{BMA} is maximum likelihood Bayesian weight.

Equation (11) can only be solved iteratively. In this study, Markov Chain Monte Carlo (MCMC) simulations based on the Differential Evolution Adaptive Metropolis (DREAM) algorithm are used to calculate the log-likelihood function. The value of β_{BMA} was used as a criterion to select better performing models that have a significant contribution in model averaging.

2.4 Integrated Bayesian Multi-model Uncertainty Estimation Framework (IBMUEF)

In this framework, the fully Bayesian approach using input uncertainty multipliers based on a specific heteroscedastic error-model as explained in section 2.2 is combined with the Bayesian Model Averaging (BMA) framework explained in section 2.3. The IBMUEF framework is implemented as follows:

- A number of alternative conceptual hydrogeological models are proposed based on
 the existing geological and hydrogeological information about the study area.
- Along with parameter uncertainty, the input uncertainty of the spatially distributed
 input data are quantified by using groundwater recharge and groundwater abstraction
 multipliers (Section 2.2 and Mustafa et al., 2018).
- 383 3. A heteroscedastic error model is defined to quantify the heteroscedasticity of thegroundwater level residual (Section 2.2).
- 4. Hydrologically reasonable prior ranges are defined for the model parameters, input
 multipliers and heteroscedastic error model parameters of each model (assuming a
 uniform prior distribution).
- 388 5. A likelihood function is defined. The likelihood function is explained in section 2.2
 389 and Mustafa et al. (2018).
- 390 6. The posterior distributions of model parameters, input multipliers and the
 391 heteroscedastic error model parameters are obtained for each model after convergence
 392 using DREAM.
- 393 7. A pre-specified number of outputs (e.g., groundwater levels) are generated for each
 394 model, using the parameter values obtained from steps 2–6.

- 395 8. The model weights and variances of each ensemble member are calculated using the396 DREAM algorithm as explained in section 2.3.
- 397 9. The model weights are computed by summing the weights for all selected ensemble398 members of each conceptual model.
- 399 10. Finally, multi-model predictions are obtained by assessing predictive mean and400 variance using equations 7 and 8.

401 **2.5 Alternative conceptual models**

Hoge et al. (2019) concluded in their review paper that selection of alternative conceptual models is the most important aspect of Bayesian Model Averaging. Enemark et al. (2019) present a review of the conceptual hydrogeological model development. In our study, four alternative conceptual groundwater flow models have been selected from 15 possible alternative conceptual groundwater flow models. These initial 15 conceptual groundwater flow models were constructed using different geological interpretations and boundary conditions.

409 All alternative conceptual models were calibrated using observed groundwater level data for 410 the same period. The performance of each model was evaluated based on different performance evaluation coefficients and information criterion statistics. Details about model 411 412 development, calibration, evaluation and selection are provided in Mustafa et al. (2019). Obviously, the best option would be to use all 15 conceptual models. However, it would be 413 414 computationally very expensive. Nevertheless, our main objective is not to predict the groundwater level of this study area. Rather our objectives are (i) to develop an integrated 415 uncertainty quantification methodology that can quantify different sources of uncertainty of a 416 417 groundwater flow model and thereby increase the reliability of the model prediction and (ii) the demonstration of the applicability of the proposed approach with real-world mode using 418 simple personal computer. Therefore, the four best performing conceptual models where 419 selected to reduce the computational effort in the Bayesian methodology. However, spatial 420 heterogeneity of the aquifer properties is not considered as a part of conceptual model 421 uncertainty. Peeters and Turnadge (2019) recommended based on their hypothetical setup 422 423 that, for an aquifer with high recharge and high conductivity, spatial heterogeneity of the aquifer properties should be considered in developing a groundwater flow model. Hence, 424 further studies could be conducted considering other alternative conceptualizations including 425 spatial heterogeneity of the aquifer properties. 426

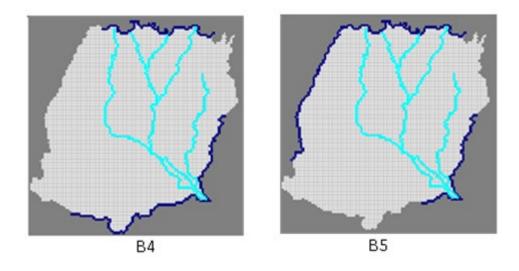
Later, the IBMUEF methodology has been implemented using the better performing four alternative conceptual models. The four selected alternative groundwater models are: (i) a one-layer model with boundary condition-5 (L1B5), (ii) a two-layer model with boundary condition-5 (L2B5), (iii) a two-layer models with boundary condition-4 (L2B4) and (iv) a three-layer models with boundary condition-5 (L3B5). Details about the selected conceptual models and model setup are explained in section 2.5.1 and 2.5.2.

433 2.5.1 Alternative conceptual models development

A cross sectional (A-A') view of the simplified hydrogeological models is shown in Figure 434 1d. First, three simplified alternative conceptual groundwater models were defined based on 435 the geological stratification. The three models are a one-layered (L1), a two-layered (L2) and 436 a three-layered (L3) model setup as shown in figure 1d. The bottom elevation of the aquifer 437 in model was taken 50 m below sea level. In the one-layered model (L1), the whole model 438 domain was considered as one hydro-stratigraphic unit and it was assumed that hydraulic 439 properties are homogeneous and isotropic. The two-layered model (L2) consists of two layers 440 where the average thickness of the top layer was 10 m (clay and loamy clay soil) and rest of 441 the thickness was considered as the bottom layer. The model domain was divided into three 442 different hydro-stratigraphic units to develop a three-layered model (L3). The top layer of the 443 three-layered model was the same as for the two-layered model, but just below the top layer, 444 a fine sand layer with an average thickness of 8 m was added in the three-layered model. The 445 446 bottom layer of three-layered model consists of medium sand, coarse sand and coarse sand with gravel. Four or more layered models were not considered in this study because the 447 448 information of the exact positions of the groundwater abstraction wells filter was unknown. Therefore, a further increase in layer numbers would increase the complexities of placing 449 450 groundwater abstraction wells in the model domain.

451 One of the major factors that influences conceptual model uncertainty is related to the boundary conditions of the model (Wu & Zeng, 2013). Boundary conditions of groundwater 452 models are often very uncertain, although the model results largely depend on these boundary 453 conditions. A previous study in the Bengal basin observed that groundwater flows from north 454 to south (Michael & Voss, 2009a, 2009b). On the other hand, there is a large wetland at the 455 southeastern corner of the study area, as well as a large river (known as Ganges/Padma) 456 457 within a few kilometers from the south boundary. Since exact boundary conditions were not known, five different potential sets of boundary conditions were conceptualized based on the 458

above information. In this study, two sets of boundary conditions are used after an initial 459 evaluation (Figure 2). Detailed description of the other boundary conditions and the 460 evaluation procedure are explained in Mustafa et al. (2019). In boundary condition 4 (B4), a 461 constant head boundary was considered on the north side of the model, where most of the 462 river branches originate, assuming that groundwater flow direction is parallel to the river 463 flow, and the southeastern part of the model, where a large wetland is located. At the south 464 part of the model domain, a constant head is assigned because the great Ganges/Padma river 465 is very near to the south boundary. In boundary condition 5 (B5), at the north and north-466 467 western boundary also at the south-eastern corner of the model domain, a constant head boundary was considered, based on the information that groundwater is flowing from north 468 and northwestern to south in the study area (Michael & Voss, 2009a, 2009b). A constant head 469 is assigned at the south-eastern corner of the model domain to represent the Chalan Beel 470 wetland. The south and north-eastern boundaries are parallel to groundwater flow direction 471 (Michael & Voss, 2009a, 2009b) hence no-flow boundaries are assigned at the south and 472 north-eastern boundaries. 473



474

Figure 2: Alternative boundary conditions used to develop alternative conceptual model (blue

476 line indicates constant head boundary): B4: constant head at north, south and southeast

477 boundary; B5: constant head at north, northwestern and southeastern boundary.

478 2.5.2 Model setup and data

PMWIN: Processing MODFLOW (Chiang & Kinzelbach, 1998) is a grid based, fullydistributed, physically-based, integrated simulation system for modelling groundwater flow
and solute transport processes and was used for groundwater flow simulations. The study area

having an area of 7112 km² was discretized into smaller cells, resulting in 117 rows and 118 482 columns of grid cells, with a dimension of 900 m x 900 m. All the alterative conceptual 483 models are transient with a monthly time step. A no-flow boundary is considered at the model 484 domain bottom as vertical groundwater flow is restricted by the relatively impermeable hard 485 rock below the aquifer in the study area. On the model top surface, a spatially distributed 486 recharge boundary is considered. Spatially distributed monthly groundwater recharge was 487 simulated using the WetSpass-M model with the same grid cell size as the MODFLOW 488 model (Abdollahi et al., 2017; Batelaan & De Smedt, 2007). The study area was divided into 489 490 34 abstraction zone considering each upazila as one zone (upazila is the second lowest tier of regional administration in Bangladesh). Groundwater abstraction in each zone was calculated 491 using an empirical relation based on the irrigated area and crop irrigation requirements. 492 Details about the estimation of the groundwater abstraction and simulation of groundwater 493 recharge can be found in Mustafa et al. (2017a). 494

495 The initial groundwater heads correspond to a long-term average groundwater table obtained496 by running the models in steady state conditions.

Weekly groundwater level and daily rainfall data were collected from the Water Resources 497 Planning Organization (WARPO), Bangladesh. The groundwater level and rainfall were 498 collected respectively for 140 and 30 sites. Available river discharge data of the BWDB for 499 the existing small rivers within the study area were also collected from WARPO. Daily 500 maximum and minimum temperature, wind speed and other climatic data were collected from 501 the Bangladesh Meteorological Department (BMD). Reference evapotranspiration (ET₀) was 502 503 calculated using the FAO Penman-Monteith equation (Allen et al., 1998; Mustafa et al., 2017a,b). In this study, reference evapotranspiration (ET₀) is also considered as potential 504 evapotranspiration. 505

- The monthly observed groundwater level data of 50 observation wells have been used formodel calibration and validation (Figure 1b).
- Topography and borehole data were collected from Bangladesh Multipurpose Development
 Authority (BMDA). The geological and lithological log data from twenty-three boreholes
 within the study area were collected from BMDA.

511 **2.6 Parameterization**

512 Groundwater recharge multipliers and groundwater abstraction multipliers have been 513 introduced to quantify uncertainty of the estimated spatially distributed groundwater recharge

and abstraction data. The input multipliers are considered as additional individual latent 514 parameters during model calibration and uncertainty analysis and have been estimated along 515 with model parameters. The hydrologically acceptable ranges of the multipliers have been 516 defined based on the available knowledge of the possible level of bias in the initial estimation 517 of groundwater recharge and abstraction (Table 1). In addition to the input multipliers, the 518 following influential MODFLOW parameters have been considered: (i) Horizontal hydraulic 519 520 conductivity, (ii) Specific yield, (iii) Hydraulic conductance of Riverbed and (iv) Specific storage. The first three MODFLOW parameters have been considered for the one-layered 521 522 model. For the two- and three-layered models, specific storage has also been added. Considering specific parameters for each layer results in, respectively, seven and ten 523 MODFLOW parameters to be considered for the two- and three-layered models (Table 1). 524 The selected parameters and their prior uncertainty ranges are presented in Table 1. 525

A uniform prior probability distribution within the hydrologically acceptable ranges has been 526 527 considered as a prior range for each parameter (Table 1) as we have no information about the distribution of the prior. Moreover, this is the most widely used prior in case of limited 528 information availability about the distribution of the parameter value (Enemark et al. 2019). 529 The range of hydrogeological parameter values was selected based on typical values for 530 aquifer materials (Domenico & Mifflin, 1965; Domenico & Schwartz, 1998; Johnson, 1967) 531 and previous research findings in the study area (Michael & Voss, 2009a, 2009b). Although 532 the number of MODFLOW parameters is different for different conceptual model structures, 533 the input multipliers and heteroscedastic error model parameters remain the same for all 534 conceptual models (Table 1). 535

Table 1: Parameters of the alternative conceptual models, input multipliers and
heteroscedastic error model parameters used in the uncertainty analysis using IBMUEF with
their prior ranges

	Descriptions	Unit	Ranges		
Input parameters for all models					
m_R	Groundwater recharge multipliers	-	0.010 - 10		
m_A	Groundwater abstraction multipliers for temporal	-	0.010 - 10		
	changes				

consider heteroscedasticity of the groundwater level error

А	Groundwater level uncertainty slope	-	0.010 - 1.0			
В	Groundwater level uncertainty intercept	m	0.010 - 3.0			
Model parameters of one-layer models (L1B5)						
HK	Horizontal hydraulic conductivity	m/s	0.0000001 - 0.0095			
RIVC	Hydraulic conductance of Riverbed	m^2/s	0.001 - 1.6			
SY	Specific yield	-	0.10 - 0.35			
Model parameters of two-layer models (L2B5, L2B4)						
HK-1	Horizontal hydraulic conductivity of layer-1	m/s	0.0000001 - 0.0095			
HK-2	Horizontal hydraulic conductivity of layer-2	m/s	0.0000001 - 0.0095			
RIVC	Hydraulic conductance of Riverbed	m^2/s	0.001 - 1.6			
SY-1	Specific yield of layer-1	-	0.10 - 0.35			
SY-2	Specific yield of layer-2	-	0.10 - 0.35			
SS-1	Specific storage multipliers of layer-1	-	0.015 - 15			
SS-2	Specific storage multipliers of layer-2	-	0.015 - 15			
Model parameters of three-layer models (L3B5)						
HK-1	Horizontal hydraulic conductivity of layer-1	m/s	0.0000001 - 0.0095			
HK-2	Horizontal hydraulic conductivity of layer-2	m/s	0.0000001 - 0.0095			
HK-3	Horizontal hydraulic conductivity of layer-3	m/s	0.0000001 - 0.0095			
RIVC	Hydraulic conductance of Riverbed	m^2/s	0.001 - 1.6			
SY-1	Specific yield of layer-1	-	0.10 - 0.35			
SY-2	Specific yield of layer-2	-	0.10 - 0.35			
SY-3	Specific yield of layer-3	-	0.10 - 0.35			
SS-1	Specific storage multipliers of layer-1	-	0.015 - 15			
SS-2	Specific storage multipliers of layer-2	-	0.015 - 15			
SS-3	Specific storage multipliers of layer-3	-	0.015 - 15			

540 2.7 Computational experiments

541 Three different scenarios have been used to perform uncertainty analysis along with model 542 calibration. The model parameters and heteroscedasticity of groundwater level error have 543 been considered in the first scenario. In this scenario, the input data are considered perfectly 544 known and accurate. This scenario will serve as a benchmark. In the second scenario, model 545 parameters, heteroscedasticity of the groundwater level error and temporal groundwater

abstraction and recharge multipliers are considered. In this scenario, we introduced 12 546 groundwater recharge multipliers (m_R) to describe uncertainties in groundwater recharge, 547 assigning a single multiplier corresponding to each time step which is one month in this 548 study. Similarly, we introduced 6 groundwater abstraction multipliers (m_A) to describe 549 uncertainties in groundwater abstraction, assigning a single multiplier corresponding to each 550 time step. Abstraction multipliers have been considered only for the dry season (November to 551 April), because this is the major abstraction period for irrigation in the study area. Details on 552 553 estimation and uncertainty analysis of groundwater recharge and abstraction can be found in Mustafa et al. (2018). 554

555 Abstraction multipliers associated with the spatial estimation have been excluded in this study because of computational time although they might have considerable effect on the 556 model prediction. In this study, four alternative conceptual groundwater models have been 557 used with different levels of complexity. The computational time increases with increased 558 complexity of the alternative conceptual groundwater models. For example, for the three-559 layer model with a total of 64 parameters (including both spatial and temporal abstraction 560 561 multipliers), the algorithm has not reached convergence even after 200000 model evaluations. On a 2.70 GHz processor, 200000 model evaluations take around 21 days with an average of 562 9 seconds per simulation. Similarly, for the two-layered model with a total of 61 parameters 563 (including both spatial and temporal abstraction multipliers), the algorithm has not been fully 564 converged after 200000 model evaluations. This corresponds with around 19 days with an 565 566 average of 8 seconds per simulation for the same processor. Of course, the evolution chain was converging towards the convergence both for the two and three-layered models. On the 567 other hand, for the one-layered model with 57 parameters (including both spatial and 568 temporal abstraction multipliers), the algorithm started to converge after 110000 model 569 evaluations. Because of time limitations, abstraction multipliers associated with the spatial 570 estimation have been excluded for all the alternative models in this study to have successful 571 convergence results for all the models. However, we believe that this will not restrict the 572 applicability of the approach because of the continuous advances in computational power. 573

574 Finally, in the third scenario, which we will refer to as IBMUEF in this study, conceptual 575 model uncertainties are considered along with uncertainties from the model input, parameters 576 and heteroscedasticity of groundwater level error. The IBMUEF framework is used to 577 quantify all the mentioned sources of uncertainty in this scenario. All the conceptual models have been calibrated and validated respectively for 1990 and 2000, for 12 monthly periods using 50 observation wells data for each period. It has been observed that models are able to accurately predict observation data which have not been used during the calibration. However, to ensure clear visualization, the results of 1990 are presented in the manuscript.

The d-factor, the % of observations within the 95 % confidence intervals (95% CI) and the
Root Mean Square Error (RMSE) have been used to evaluate the model prediction
uncertainty. The d-factor represents the average width of the 95% CI and is calculated as in
(Yang et al., 2008):

$$d - factor = \frac{\frac{1}{n}\sum_{t=1}^{n}(H_{t,u} - H_{t,l})}{\sigma_0}$$
(12)

587 Where $H_{t,u}$ and $H_{t,l}$ represent respectively, the upper and lower bounds of the 95% confidence 588 intervals, n = the number of observations and σ_0 = the standard deviation of the observed 589 groundwater level. d-factors closer to 1 indicate better model prediction (Yang et al., 2008). 590 The higher observation coverage within the 95 % confidence intervals and decreasing d-591 factor value are indicating the improvement in model predictions and accuracy of the 592 uncertainty bounds.

593 3. Results and discussion

594 In the results and discussion section, the results obtained from the three different scenarios as 595 explained in the previous section (section 2.7) are presented, interpreted and discussed. Section 3.1 presents the parameter and prediction uncertainty of different conceptual models 596 597 due to uncertainty of model parameters along with the heteroscedastic error model parameters. Section 3.2 elaborates on the parameter and prediction uncertainty of different 598 599 conceptual models due to the uncertain input, model parameters along with the 600 heteroscedastic error model parameters. Finally, section 3.3 presents the prediction 601 uncertainty due to uncertainty of the conceptual model structure, input, model parameters and parameters of the heteroscedastic error model. 602

603 **3.1** Parameter and prediction uncertainty of different conceptual models for scenario 1

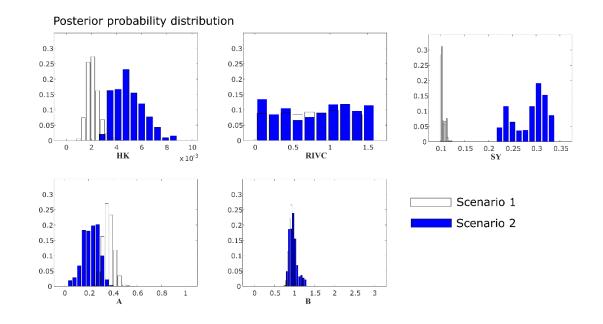
Figure 3 shows the posterior probability distributions of the L1B5 model parameters for scenario 1. All parameters except riverbed hydraulic conductance (RIVC) of L1B5 model are well identified within their prior distribution. The posterior distribution of RIVC is still

almost uniform while the posterior distribution of all other parameters is normally distributed, 607 indicating that RIVC is a non-influential parameter. However, this could be improved in 608 future studies by including more streamflow data during model calibration. We have also 609 examined the correlation between model parameters and error model parameters. The results 610 show a weak correlation among the MODFLOW parameters and between model parameters 611 and error model parameters. The posterior distribution of SY is located at the lower 612 boundaries of the prior range with a mean value of around 0.11. Alternatively, the posterior 613 distribution of horizontal hydraulic conductivity (HK) is almost normally distributed with a 614 high mean value of around 2.5 x 10^{-3} ms⁻¹. However, different conceptual models with 615 different parameterization might draw different conclusions. Hence, consideration of 616 conceptual model structural uncertainties may be important, but this is not considered in this 617 scenario. Although the posterior probability distributions of the well identified parameters 618 cover only a small range of their prior distributions, the parameter uncertainty band covers 619 only 8.5% of the observations (Figure 5a). This can be argued as a problem of 620 overconfidence in the estimation of the model parameters. Though the total uncertainty band 621 covers almost all observations (94%), the width of the total uncertainty band is very wide 622 compared to the width of the parameter uncertainty band. This is indicating that both the 623 considered conceptual model structure and the input data used for this scenario contain a 624 considerable amount of uncertainty. 625

Figure 4 shows the posterior pdfs of the L3B5 model parameters for scenario 1. As expected, 626 the posterior parameter distributions of the L3B5 model are very different from the posterior 627 parameter distributions of the L1B5 model. In this scenario, 12 parameters are considered, 628 including two parameters of the heteroscedastic error model (A and B). Out of these 12 629 parameters, the posterior distributions of six parameters (HK-1, HK-2, HK-3, SY-1, a, and b) 630 631 are approximately normally distributed. The posterior distribution of riverbed hydraulic conductance (RIVC) is still almost uniform like its prior distribution, again indicating that 632 RIVC is a non-influential parameter. The posterior distributions of specific storage 1, 2 and 3 633 (SS-1, SS-2 and SS-3) are not included in the figure as the posterior distributions of those 634 parameters are also still almost uniform as were their prior distributions. Similarly, the 635 posterior distributions of specific storage for the two layered models also remain uniform, 636 indicating that this is also a non-influential parameter (supplementary materials: 637 Supplementary Figure 1). The posterior distributions of HK-1 and SY-2 are located 638 respectively at the lower and upper boundaries of the prior range. Moreover, the posterior 639

distribution of SY-3 is not well identified. This could be due to input uncertainties and/or 640 conceptual model structural uncertainties which are not considered in this scenario. It also 641 shows that the posterior probability distributions of the well identified parameters cover only 642 a small range of their prior distributions except for HK-2. The parameter uncertainty band 643 covers only 13 % of the observations (Figure 5d). Similar results are observed for the L2B4 644 and L2B5 models. For the L2B4 and L2B5 models, the parameter uncertainty band covers 645 respectively 12 % and 13.8 % of the observations (Figure 5b, 5c and Supplementary Table 1). 646 In general, the parameter uncertainty band is increasing with the level of complexity of the 647 648 conceptual models. The observation coverage of the parameter uncertainty band for the different conceptual model structures is different. This suggests the importance of the use of 649 multiple conceptual models for reliable prediction. Hoge et al. (2019) also suggested that 650 consideration of uncertainty arising from conceptual physical interpretation is important 651 during BMA implementation, if the objective of the study is to increase the reliability and 652 accuracy of the model prediction. 653

654



655

Figure 3: The posterior probability distribution of the L1B5 model parameters (top row) and
the parameters of the heteroscedastic error-model (bottom row) both for scenario 1 and 2,
using 2500 samples generated after convergence. HK: Horizontal hydraulic conductivity,
RIVC: Hydraulic conductance of riverbed, SY: Specific yield, A and B: The parameters of
the heteroscedastic error model.

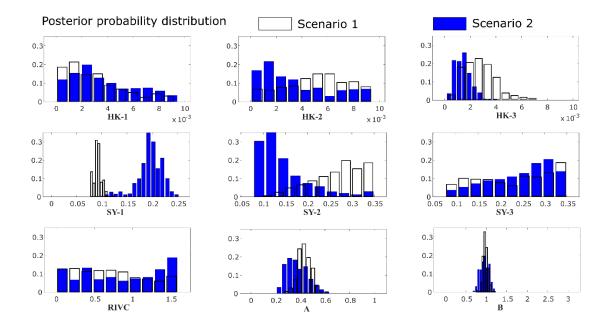


Figure 4: The posterior probability distribution of the L3B5 model parameters and the parameters of the heteroscedastic error-model (A and B) both for scenario 1 and 2, using 2500 samples generated after convergence.

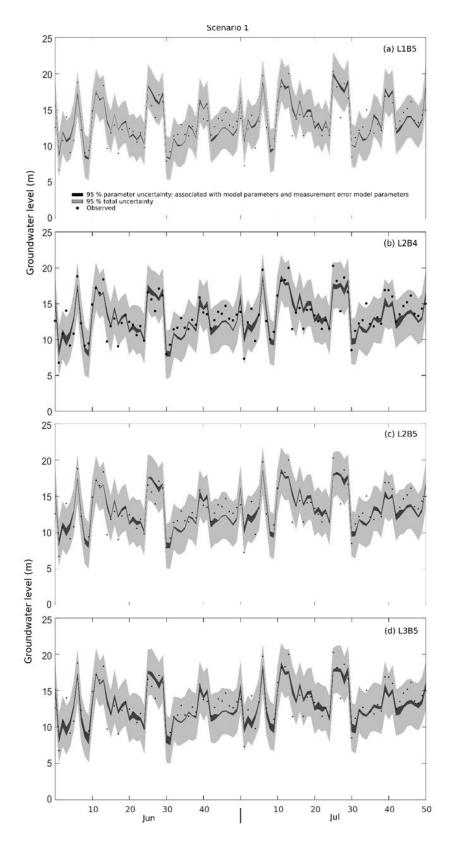


Figure 5: The prediction uncertainty of monthly groundwater level at each observation well
with 95% parameter uncertainty considering error-model parameter along with model
parameter (black interval), 95 % total uncertainty (dark gray) and observations (black dot) for
(a) L1B5 model, (b) L2B4 model, (c) L2B5 model and (d) L3B5 model.

672 **3.2** Parameter and prediction uncertainty of different conceptual models for scenario 2

In this scenario, uncertainty of the input data is quantified simultaneously along with modelparameters and heteroscedastic error-model parameters.

Figure 3 shows the posterior pdfs of the L1B5 model parameters for scenario 2. As in 675 676 scenario 1, all parameters are well identified within their prior ranges except RIVC. The posterior pdfs of the well identified parameters cover only a limited part of the prior range. 677 The posterior distribution of the hydraulic conductance of riverbed (RIVC) is still almost 678 uniform. Additionally, the posterior distribution of SY shows a slight multimodality. The 679 correlation among model parameters and the correlation between model parameters, error 680 model parameters and input multipliers have been examined. The results show a weak 681 correlation among the MODFLOW parameters and between model parameters, error model 682 parameters and input multipliers (recharge and abstraction multipliers). 683

684 Out of the 12 parameters for model L3B5, the posterior distributions of eight parameters 685 (HK-1, HK-2, HK-3, SY-1, SY-2, SY-3, a, and b) are approximately normal while it was six 686 for scenario 1 (Figure 4). The posterior distribution of RIVC, SS-1, SS-2, SS-3 are still 687 almost uniform.

By comparing the posterior distributions between scenario 1 and 2 for different conceptual
models (Figures 3 and 4), the following observations are made:

- The posterior pdfs of some parameters are different in different conceptual models as
 well as in different scenarios. This is indicating that parameter values are overly
 adjusted to compensate for existing conceptual model structural deficiencies and input
 uncertainty when input and/or conceptual model uncertainties are not considered.
- 694
 2. For model L3B5, the posterior pdfs of the parameters SY-2 and SY-3 are also
 695 identified within their prior ranges and their posterior distribution became
 696 approximately normal when we consider input uncertainty in addition to uncertainty
 697 arising from model parameters and heteroscedastic error model parameters. However,
 698 their posterior distributions are located at the boundaries of the prior range. This could
 699 be because of model structural uncertainties.
- 3. The heteroscedastic error model parameters (A and B) are well identified in bothscenarios for all different conceptual models, but their values are different between

scenarios and between models. In general, the values of the error heteroscedasticity
(A and B) parameters decrease when we consider input uncertainty in addition to
uncertainty of model parameter and heteroscedastic error model parameters. Another
important observation is that the value of the first error heteroscedasticity (A)
parameter increases with the level of complexity of the conceptual models. This
indicates that existing conceptual model structural deficiencies are somehow
compensated by the value of the error heteroscedasticity (a) parameter.

We conclude that an explicit consideration of input uncertainty in addition to uncertainty of the model parameters and heteroscedastic error model parameters is very important to have unbiased and better defined parameter sets. Consideration of alternative conceptual models is also important for obtaining confident parameter sets. Schoniger et al. (2015) also reported that consideration of uncertainty arising from the model input is necessary to increase the robustness of Bayesian model averaging and ranking.

715 The posterior probability distributions of the recharge multipliers vary strongly between months, but are in general higher than one. The recharge multipliers are well identified during 716 the rainy season (May to October), while these multipliers are not well identifiable during the 717 dry season (November to April). The details of the recharge multipliers for a specific 718 719 conceptual model are explained in Mustafa et al. (2018). The distributions of the well identified multipliers show different shapes for different conceptual models (Figure 6). 720 However, the range of the multipliers and magnitude of their probability distributions are the 721 same for different conceptual models (Figure 6). The groundwater abstraction multipliers are 722 723 also well identified within their prior range and are higher than one in all months except for November and January for all four conceptual models. Again, the well identified multipliers 724 725 show almost the same range of values for different conceptual models (Figure 7). This 726 indicates that the input uncertainty multipliers are independent from model structural uncertainty and are not overly adjusted to compensate conceptual model structural 727 deficiencies. 728

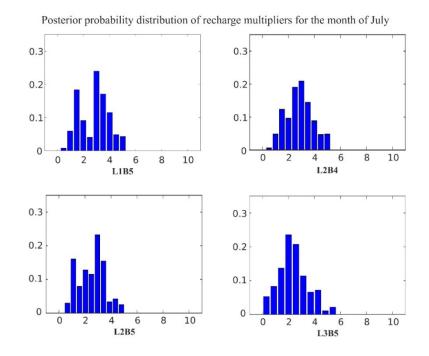
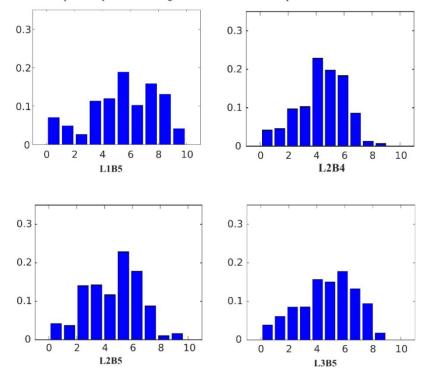


Figure 6: Posterior distribution of groundwater recharge multipliers of July for all conceptual

731 models, using 2500 samples generated after convergence.



Posterior probability distribution of groundwater abstraction multipliers for the month of December

Figure 7: Posterior distribution of groundwater abstraction multipliers, using 2500 samplesgenerated after convergence.

The prediction uncertainty of the monthly groundwater level associated with input

uncertainty, model parameter uncertainty and uncertainty related to the heteroscedastic error 737 model is presented in figure 8. The observation coverage of the parameter uncertainty band 738 has increased by more than 100% for all models (Supplementary Table 1) when uncertainty 739 arising from model input is incorporated along with uncertainty arising from model 740 parameters and parameters of the heteroscedastic error model. The increase for the L1B5 741 model is even more than 200%. This result reveals that consideration of input uncertainty has 742 significantly improved the confidence of model predictions and ignoring input uncertainty 743 could lead to biased model simulations and incorrect uncertainty bands.. The parameter 744 uncertainty band of L1B5 covers the highest number of observations when input uncertainty 745 746 is included (Supplementary Table 1). When we explicitly consider input uncertainty, the width of the parameter uncertainty band has increased but the width of the total uncertainty 747 748 has decreased (figure 5 and 8). This indicates that total uncertainty has decreased. This is confirmed by the reduction of the d-factor for all the models (Supplementary Table 1). This 749 result reveals that uncertainty bounds of scenario 2 are more accurate compared to the CI of 750 scenario 1, and the residual variance is smaller at each point. The Root Mean Square Error 751 (RMSE) was also used to compare the results of scenario 1 and 2. It is observed that the 752 values of the RMSE are decreasing when input uncertainty is included along with model 753 parameter uncertainty and the parameters of the heteroscedastic error model (Figure 14). The 754 decreasing magnitude of the RMSE value of L1B5 model is more significant than for any of 755 756 the other models, indicating comparatively higher uncertainties in the L1B5 model. This is another indication that consideration of uncertainties through input multipliers is increasing 757 758 the accuracy of the model prediction and decreasing the prediction uncertainty. Even after 759 consideration of input uncertainties, the observation coverage of the parameter uncertainty 760 band for the different conceptual model structures is different (Supplementary Table 1, Figure 8). Hence, consideration of conceptual model structural and input uncertainty is 761 762 important to have more accurate model prediction and unbiased uncertainty bounds.

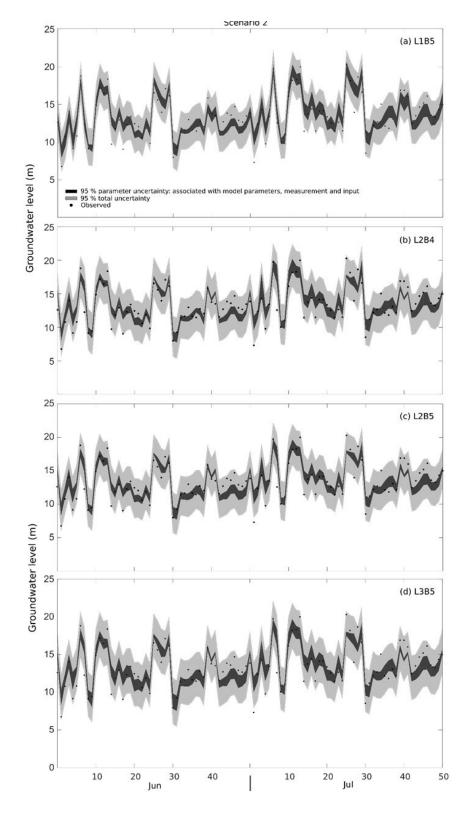




Figure 8: Prediction uncertainty of monthly groundwater level at each observation well with
95% parameter uncertainty considering input uncertainty along with model parameter
uncertainty and error heteroscedasticity (black interval), 95 % total uncertainty (dark gray)
and observation (black dot) for (a) L1B5 model, (b) L2B4 model, (c) L2B5 model and (d)
L3B5 model.

3.3 Application of IBMUEF: assessment of the model uncertainty from input, model parameters, parameters of the heteroscedastic error-model and conceptual model structure

In the IBMUEF framework, uncertainties originating from the model input, the parameters, the parameters of the heteroscedastic error-model and the conceptual model structure can be taken into account. In this section, besides a presentation and discussion of the results of the IBMUEF application, these are also compared with the results of the previous scenarios.

Figure 9 shows parameter uncertainty bounds for all four alternatives conceptual models considering uncertainty arising from model input, parameter, and measurement heteroscedasticity. The different conceptual model structures cover different observations. This is indicating the skill of the models to capture different hydrogeological processes of the system.

The marginal densities of the estimated weights (following step 9 of section 2.4) for each 781 model are shown in figure 10. The weight of the L1B5 model is well identified and has a 782 normal distribution. Its likelihood value (weight) is very high compared to other models. The 783 weight of all other models is very small. Nevertheless, their contribution is considered in the 784 final results as they are representing different geological processes which are not considered 785 in L1B5. The necessity of incorporating different models is also confirmed by the limited 786 correlations between the groundwater level predictions using different conceptual models 787 (Supplementary Table 2). For example, if a researcher/user knows the L1B5 model 788 prediction, the L2B5 model adds more additional information to the final result compared to 789 790 the L2B4 and L3B5 models as L2B5 is less correlated with L1B5. In general, correlations between the models are limited, indicating that different conceptual models are providing 791 important information of the different hydrogeologic processes of the system. Hence, 792 consideration of different conceptual models is needed to have a reliable model prediction. 793

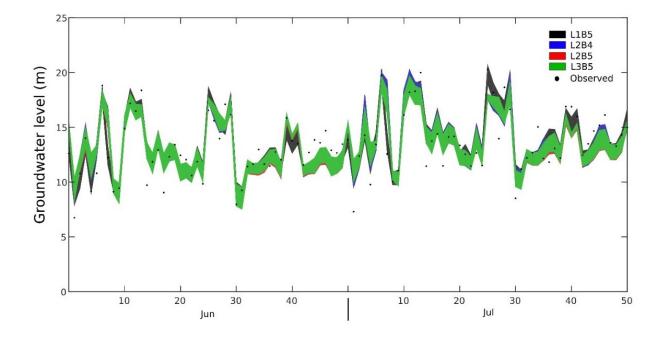




Figure 9: Prediction uncertainty of monthly groundwater level for all the conceptual models
at each observation well with 95% parameter uncertainty considering input uncertainty along
with model parameter uncertainty and error heteroscedasticity and observation (black dot).

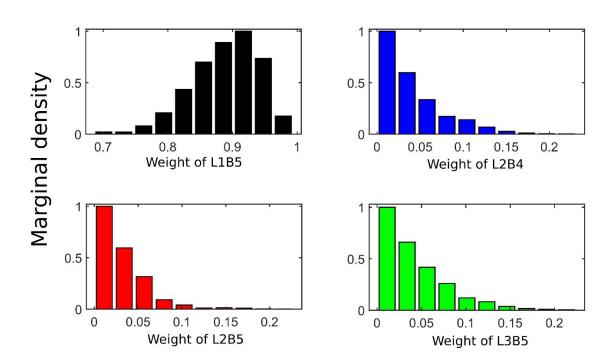


Figure 10: Marginal density of estimated weight for each model using integrated Bayesianmulti-model uncertainty estimation framework (IBMUEF).

Figure 11 shows the final IBMUEF 95% prediction uncertainty of the monthly groundwater 805 levels at each observation well considering model input, parameter, error heteroscedasticity 806 model parameter, and conceptual model structural uncertainty. The final IBMUEF prediction 807 was calculated using the prediction of the individual member models and their corresponding 808 likelihood values (Figures 9 and 10) as explained in sections 2.3 and 2.4. The black line in 809 figure 11 shows the mean prediction of the IBMUEF. The IBMUEF mean prediction and 810 variance of figure 11 were calculated using equation 7 and 8, respectively. The distribution 811 812 shape is determined by the weighted sum of the posterior distributions of each member model (Figure 12). It is observed that the posterior distribution of L1B5 model is capturing the 813 reality more accurately compared to other models in the selected section of figure 11 (Figure 814 12). Hence, the distribution shape of the L1B5 model has a dominant role on the final 815 816 IBMUEF prediction distribution shape of that section.

As expected, the 95 % CI of IBMUEF covers 95% of the observations which is very high 817 compared to the individual models (Figure 11 and 13). Another interesting observation is that 818 the d-factor value (1.42) has become smaller than the previous results. This is an indication of 819 the improved model predictions and accuracy of the uncertainty bounds. The Root Mean 820 Square Error (RMSE) was also used to evaluate the skill of the IBMUEF and to compare it 821 with the individual model ensembles. The probability distributions of the RMSE-values for 822 each of the models and IBMUEF are shown in figure 14. It is observed that the IBMUEF 823 824 results in lower RMSE values compared to any individual model from the ensemble (Figure 14). This result reveals that the IBMUEF framework provides better model predictions. We 825 conclude that an explicit consideration of conceptual model structural uncertainty is 826 827 important for obtaining more accurate model predictions and unbiased uncertainty bounds. The results for this study are in line with results from similar approaches in surface water 828 829 modeling (e.g., Ajami et al., 2007).

The IBMUEF framework is providing better and more reliable model predictions and more accurate uncertainty bounds, which is very important for decision support applications. However, as mentioned earlier, the implementation of the methodology is computationally expensive. The computational burden has also been identified as a main drawback for all other existing integrated uncertainty assessment approaches (Rojas et al., 2008; Ajami et al., 2007; Gelman et al., 2014). Based on their hypothetical setup, Xue & Zhang (2014) and

Hendricks Franssen et al. (2011) advocated that the EnKF is computationally more efficient 836 compared to other existing Bayesian methods. However, comparison of the computational 837 efficiency of the EnKF and other integrated Bayesian approaches with a real-world model 838 remain unsolved. Another alternative could be Granger-Ramanathan averaging (GRA). GRA 839 provides very similar performance as BMA, but is computationally less demanding (Diks and 840 Vrugt, 2010). The information criterion (e. g.: AIC: Akaike information criterion) is another 841 alternative to obtain a computationally less demanding approach (Hoge et al. 2019). 842 However, model averaging based on AIC has been criticised by researchers as it is not based 843 844 on a rigorous statistical basis and its results has no BMA interpretation (Wasserman, 2000; Tsai and Elshall, 2013). That's why it has been considered as a model selection technique 845 instead of model averaging (Hoge et al. 2019). As a consequence, we have to choose between 846 two different approaches: (i) computationally demanding but statistically robust, reliable and 847 more accurate approaches or (ii) approaches without rigorous statistical foundation, which are 848 computationally less demanding. Nonetheless, the statistically robust adaptive MCMC 849 sampling of the DREAM algorithm is computationally more efficient for high-dimensional 850 851 and multimodal application (Vrugt et al 2009a, 2016; Leta et al., 2015). Since the latter multichain MCMC based algorithm has been adopted in this study as a sampling approach, we 852 853 believe that our approach is computationally more efficient compared to other existing integrated uncertainty assessment approaches. Moreover, the IBMUEF is a flexible 854 855 framework as (i) there is no limitation for the number or complexity of alternative conceptual models and (ii) users can choose the number and dimensions (spatial and temporal) of input 856 multipliers, based on the objectives of their modelling. It should be remembered that, the 857 computational time increases with increases complexity of the alternative conceptual 858 groundwater models and for a very complex model with more than 60 model parameters, the 859 proposed approach became computationally very expensive. However, we believe that this 860 861 will not restrict the applicability of the approach because of the continuous advances in computational power. Even though, effort should continue in the development of a more 862 computationally efficient approach. We conclude that number or complexity of alternative 863 conceptual models should be considered based on the modelling objectives during the 864 implementation of a integrated uncertainty assessment approaches. Hoge et al. (2019) also 865 concluded that the objective of the modelling should be the main driver in selecting model 866 averaging approaches. 867

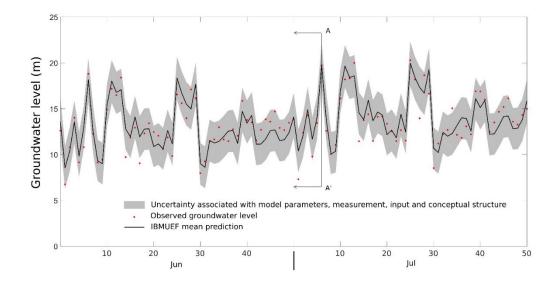


Figure 11: 95 % prediction uncertainty of monthly groundwater level at each observation well considering model input, parameter, error heteroscedasticity model parameter, and conceptual model structural uncertainty (gray shad), and IBMUEF predictive mean (black line), observation (red dot).

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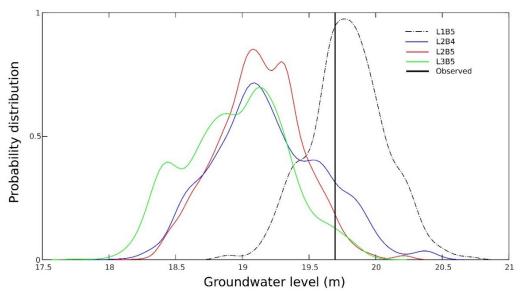
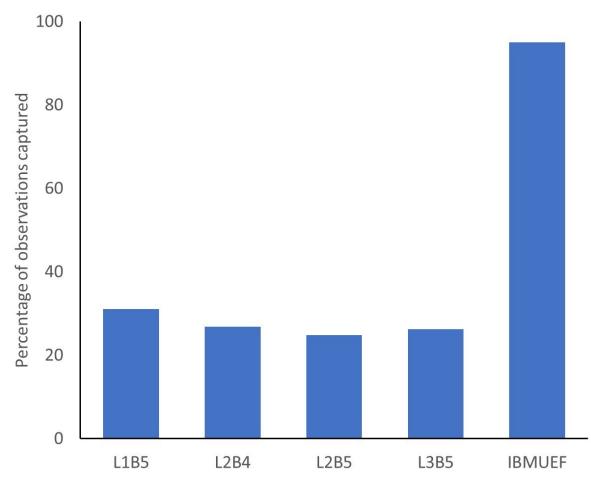




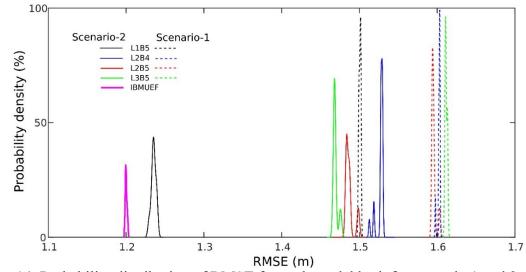
Figure 12: Posterior probability distribution of groundwater level prediction for each member model at the selected cross-section (A-A') of figure 11 and observed groundwater level (black line).

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879
880 Figure 13: Percentage of observation captured by 95% parameter uncertainty bands of each
881 conceptual model and IBMUEF.

882



RMSE (m)
Figure 14: Probability distribution of RMSE for each model both for scenario 1 and 2 and
IBMUEF.

886 **4.** Conclusions

We present an integrated Bayesian multi-model uncertainty estimation framework (IBMUEF) 887 to explicitly quantify the uncertainty originating from errors in model conceptualization, 888 input data, parameter values and measurement heteroscedasticity error of a fully distributed 889 physically-based groundwater flow model. In the proposed integrated fully Bayesian multi-890 model framework, the DREAM algorithm with a specific likelihood function is combined 891 with BMA. Groundwater recharge multipliers and groundwater abstraction multipliers are 892 used in this framework to quantify uncertainty of spatially distributed input data of the 893 groundwater model. The measurement heteroscedasticity is also considered in our integrated 894 895 Bayesian framework by incorporating a novel heteroscedastic error model. To check the applicability of IBMUEF, four alternative conceptual models have been developed using a 896 897 numerical groundwater flow model (MODFLOW) based on different interpretations of geological and hydrogeological information about the study area. 898

The results of this study confirm that conceptual model structure and uncertainty on the input 899 data have a considerable effect on the model parameter distributions and model predictions. 900 We demonstrated that parameter values are overly adjusted to compensate the existing 901 conceptual model structural deficiencies and input uncertainty when they are not taken into 902 account. Although consideration of input uncertainty results in better defined parameter 903 distributions, consideration of alternatives conceptual models is also important to obtain 904 confident parameter sets as the existing conceptual model structural deficiencies are 905 906 somehow compensated by parameter uncertainties and the parameters of the heteroscedastic 907 error model. On the other hand, input uncertainty multipliers appear to be independent from 908 model structural uncertainty.

909 The total uncertainty of the system decreases but the observation coverage of the parameter 910 uncertainty band increases by more than 100 % for all considered models when input uncertainty is included. Even when considering input uncertainty, the observation coverage 911 of the parameter uncertainty band for the different conceptual model structures is different. 912 This suggests the importance of the use of multiple conceptual models for reliable prediction. 913 The parameter uncertainty band of L1B5 covers the highest number of observations when 914 input uncertainty is included. This indicates that the L1B5 model is more capable of 915 capturing the reality when input uncertainty is included. This is also confirmed by the highest 916 likelihood (weight) value of the model. We demonstrate that consideration of input 917

918 uncertainty along with model parameters uncertainty and measurement error generate more 919 reliable model predictions. However, a very common limitation of these results is that the 920 results are based on only a single conceptual model. Our results also confirm that even a very 921 well calibrated conceptual model is unable to represent all the hydrogeologic processes of the 922 system.

The IBMUEF prediction was calculated using the prediction of the individual member 923 models and their corresponding likelihood values. The 95% prediction uncertainty band of 924 IBMUEF covers 95 % of the observations which is significantly higher compared to any of 925 the individual models. The IBMUEF framework has decreased the RMSE-value of the 926 927 prediction and d-factor of the CI, and thereby increased the reliability of the prediction. The results of the study confirm that the IBMUEF framework is a useful tool to have better and 928 929 more reliable model predictions and accurate uncertainty bounds. It is also shown that the IBMUEF is a useful and applicable framework to simultaneously quantify input, parameter, 930 931 measurement and conceptual model uncertainty of a fully distributed physically-based groundwater flow model. We conclude that an explicit consideration of conceptual model 932 structural uncertainty along with model input, parameter and measurement uncertainty using 933 IBMUEF framework improves the accuracy and reliability of the model prediction and 934 related uncertainty bounds. 935

936 Alternative conceptual models considered in this study have been developed using only 937 MODFLOW. Future studies could be conducted considering different groundwater modelling 938 algorithms to quantify the effect of numerical modelling errors. The modified log-likelihood 939 function as explained in section 2.2, has been used in this study. However, future studies 940 could be conducted considering different likelihood function to evaluate the effect of 941 likelihood function.

In future studies, the framework can be implemented with more additional data sets to check the applicability with different prediction objectives e.g: baseflow. Moreover, application of the IBMUEF framework to quantify the groundwater level prediction uncertainties originating from the climate change and abstraction scenarios will increase the reliability of the model prediction and accuracy of the uncertainty bounds as its (IBMUEF) already consider all the other sources of uncertainties. However, number or complexity of alternative conceptual models or future scenarios should be considered based on the modelling 949 objectives during implementation of this integrated uncertainty assessment approaches to950 avoid conceptual burden.

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AgEng2018.com but has not been published. The data used in this study are summarized and

presented in the figures, tables, references and supporting information and will be availablefrom the authors upon request (syed.mustafa@vub.be).

957 Author contributions

SM, JN and MH designed the study. SM and GG performed the analysis. SM and MH wrotethe manuscript. All authors discussed the results and commented on the manuscript.

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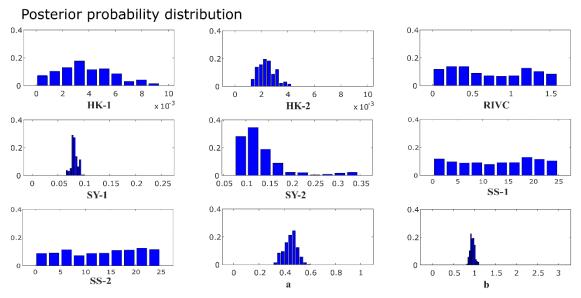
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1172 Supplementary material



1173
1174 Supplementary Figure 2. The posterior probability distribution of the L2B4 model
1175 parameters and the parameters of the heteroscedastic error-model (A and B) for scenario 1,

using 2500 samples generated after convergence.

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- **Supplementary Table 3.** Percentage observation coverage of the parameter uncertainty band
- and calculated d-factor based on the total uncertainty band for all the conceptual models.

	L1B5		L2B4		L2B5		L3B5	
	% cover	d-factor						
Scenario 1	8.5	1.88	12.0	2.03	13.8	2.01	13.0	2.04
Scenario 2	31.0	1.59	26.8	1.94	24.8	1.89	26.16	1.88

- **Supplementary Table 2.** Correlations between the groundwater level predictions using
- 1181 different conceptual models.

	L1B5	L2B4	L2B5	L3B5
L1B5	1	-0.495	-0.367	-0.491
L2B4		1	-0.125	-0.119
L2B5			1	-0.072
L3B5				1